

VISIBLE-LIGHT PHOTOELECTRODEGRADATION OF DIURON ON WO₃ NANOSTRUCTURES

Fernández-Domene, R.M.; Sánchez-Tovar, R.; Lucas-Granados, B.; Muñoz-Portero, M.J.; Ramírez-Grau, R.; García-Antón, J.*

Ingeniería Electroquímica y Corrosión (IEC). Departamento de Ingeniería Química y Nuclear. ETSI Industriales. Universitat Politècnica de València. Camino de Vera s/n, 46022 Valencia, Spain. Tel. 34-96-387 76 32, Fax. 34-96-387 76 39,

*e-mail: jgarciaa@iqn.upv.es (J. García-Antón).

The degradation of pesticide diuron has been explored by photoelectrocatalysis (PEC) under visible light illumination using two different WO₃ nanostructures, obtained by anodization of tungsten. The highest degradation efficiency (73%) was obtained for WO₃ nanosheets synthesized in the presence of small amounts of hydrogen peroxide (0.05 M). For that nanostructure, the kinetic coefficient for diuron degradation was 133% higher than that for the other nanostructure (anodized in the presence of fluoride anions). These results have been explained by taking into account the different architecture and dimensions of the two WO₃ nanostructures under study.

KEYWORDS: Photoelectrocatalysis; WO₃ nanostructures; anodization; diuron; pesticides.

1. INTRODUCTION

Photoelectrocatalysis (PEC) is an advanced oxidation process which has been investigated as an efficient method for the removal of recalcitrant organic molecules from wastewaters [1-4]. PEC is a multidisciplinary field, involving electrochemistry, materials science, solid-state physics and optics [3]. When light of enough energy reaches the surface of a semiconductor, electrons from the valence band are excited to the conduction band, leaving a positive-charged hole behind, hence creating photogenerated charge carriers, i.e., electron-hole pairs ($e^- - h^+$). Unlike simple photocatalysis (PC), in PEC a small potential is applied to the photocatalyst (*n*-type semiconductor) to minimize the recombination of the photogenerated charge carriers (see scheme in **Figure 1**). This external polarization increases the semiconductor band bending and sustains the electric field inside the space-charge region, thus favoring the transport of electrons from the semiconductor conduction band to the counter electrode and increasing the lifetime of holes [2]. Valence-band holes oxidize the aqueous medium to generate hydroxyl radicals, which are extremely oxidant species and can react with the recalcitrant organic molecules (R in **Figure 1**). Additional and more complex configurations for PEC cells can be found in the literature [5-8].

Nanostructured semiconductors are ideal for PEC, since their high surface area facilitates the interaction between the semiconductor surface and the electrolyte, creating many active sites where photogenerated holes can react with water molecules to generate hydroxyl radicals. Besides, charge transfer and light absorption are highly enhanced by using nanostructures [3, 9]. Among the semiconductor oxides investigated for PEC applications, WO_3 nanostructures present interesting properties, such as band

gap values around 2.6 eV ($\lambda = 477$ nm) that permit the absorption of visible light and high chemical and photoelectrochemical resistance in acidic environments [1, 10]. WO_3 photoanodes have been synthesized to degrade organic compounds such as dyes [11-21], pesticides [22] or drugs and other chemicals [23-27].

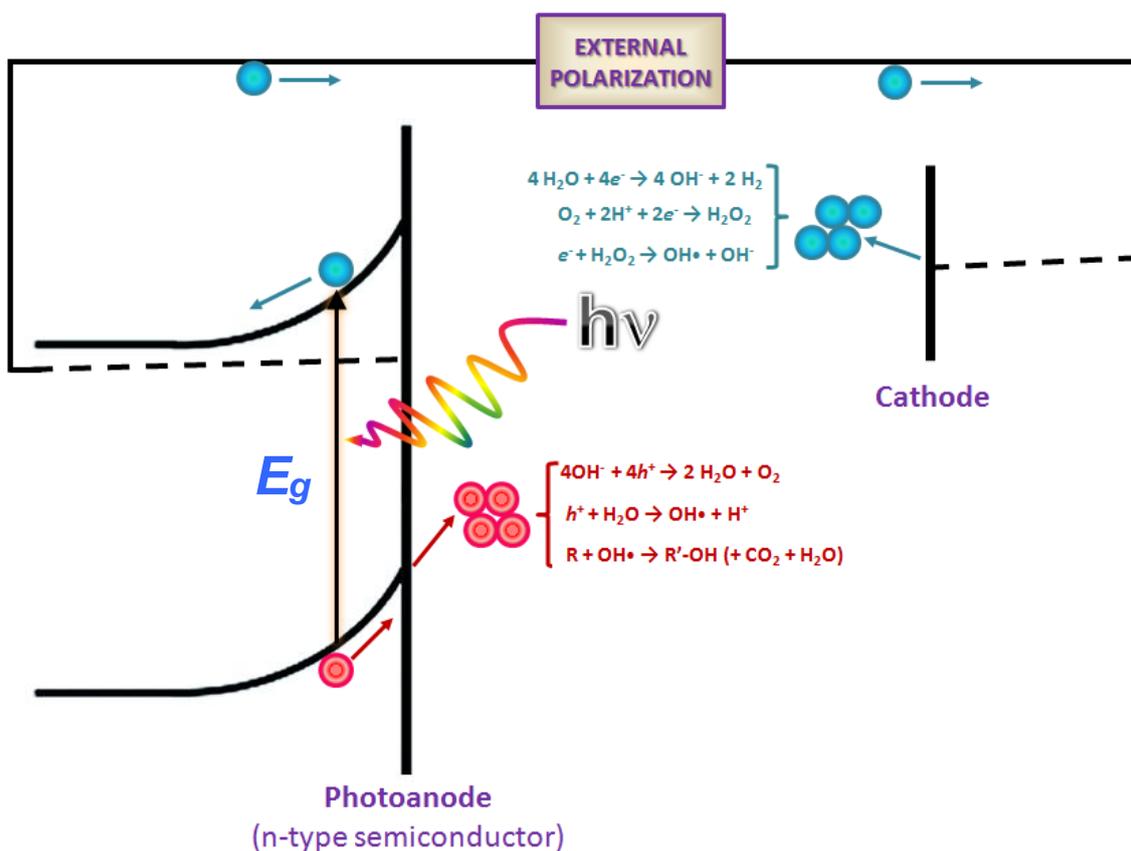


Figure 1. Scheme of a photoelectrocatalytic cell.

This work studies the PEC degradation of diuron (**Figure 2**), a persistent herbicide, by using high-performance WO_3 nanostructures. Diuron (3-(3,4-dichlorophenyl)-1,1-dimethylurea) is an inhibitor of photosynthesis broadly used in both crop and non-crop areas. It is a substance very toxic to aquatic life and it is suspected of causing cancer and damage to organs through prolonged or repeated exposure [28-36]. Diuron is persistent, mobile and has been found in surface and groundwater [29, 37-49]. In this work, an

effective and feasible PEC method for diuron elimination from wastewaters is presented for the first time.

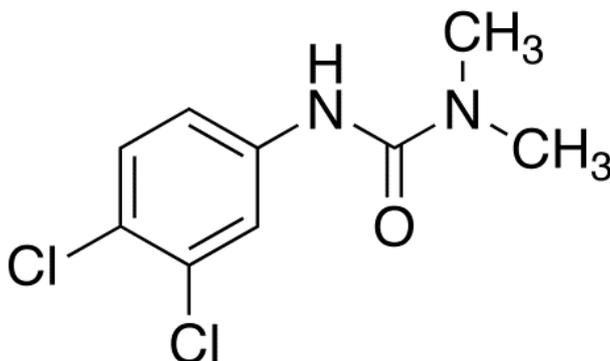


Figure 2. Diuron molecule.

2. MATERIALS AND METHODS

Aqueous solutions of diuron were prepared from analytical standard grade diuron supplied by Sigma-Aldrich (Pestanal[®]). Solutions were prepared with distilled water and stirring for 24 h in the dark.

WO₃ photoanodes were prepared by electrochemical anodization of tungsten rods (8 mm in diameter) immersed in two different electrolytes: 1.5M H₂SO₄ + 0.1M NaF and 1.5M H₂SO₄ + 0.05M H₂O₂, both at 50° C. Tungsten rods were covered with a Teflon coating to prevent the rest of the surface from contacting the electrolyte (0.5 cm² of surface exposed to the solution). The cell potential difference between the tungsten working electrode and a platinum mesh counter electrode was set at 20 V for 4h, and the tungsten electrode was continuously rotated at 375 rpm by using a rotating disk electrode (RDE), according to a previous work [50]. After anodization, samples were annealed in a tubular furnace at 400° C for 4h in an air atmosphere.

WO₃ nanostructures were morphologically characterized by using a Field Emission Scanning Electron Microscopy (FESEM).

PEC performance of the WO₃ nanostructures was evaluated by degrading a 20 mg L⁻¹ diuron solution with 0.1 M H₂SO₄ as a supporting electrolyte, under 420 nm light (by using a 1000 W Xe light source, 100 mW cm⁻² at 420 nm) and applying an external potential of 1 V. The acidic supporting electrolyte was used to enhance the photoelectrocatalytic performance of nanostructures, according to a previous work [21]. The reactor was a quartz glass cell with three electrodes: the WO₃ nanostructure was the working electrode (0.5 cm² exposed to the diuron solution), an Ag/AgCl (3M KCl) was the reference electrode and a Pt tip was the counter electrode. In total, 14 ml of the diuron solution were added to the reactor. Prior to the degradation tests, samples were immersed in the diuron solution for 30 minutes at their open circuit potential to achieve initial equilibrium conditions. Photocurrent density values were recorded during the whole tests.

At regular time intervals of the reaction (every 30 minutes), 4 mL of the diuron solution were taken and analyzed immediately by UV-visible spectrophotometry, measuring the absorbance between 190 nm and 300 nm. After that, the aliquot was returned to the reactor to continue with the process. The extent of the degradation reaction was also quantified by measuring the concentration of chloride and nitrate anions resulting from the decomposition of the diuron molecule (see **Figure 2**). This measure was performed with a Metrohm 883 Basic IC plus ion chromatograph equipped with a Metrosep A Supp 5⁻ (150/4 mm) separation column. The eluent was a carbonate/bicarbonate (3.2/1 mM) buffer solution at a flow rate of 0.7 mL min⁻¹ and the suppressor was a 0.1 M

H₂SO₄ solution. Under these experimental conditions, retention times for both anions were ~6 min for Cl⁻ and ~10.3 min for NO₃⁻.

3. RESULTS AND DISCUSSION

3.1. Nanostructures morphology: influence of anodization electrolyte

The results of the morphological FESEM analysis of the WO₃ nanostructures are presented in **Figure 3**. The most characteristic feature of the nanostructure fabricated in the NaF electrolyte was the formation of ordered nanoplatelets arranged in 3D globular clusters (**Figure 3a**). The length of these nanoplatelets was determined to be 626 ± 25 nm, their width 29.2 ± 1.2 nm, and the thickness of the whole nanoplatelet layer 15.5 ± 1.5 μm. When synthesized in the presence of H₂O₂ (instead of NaF), a morphological change in the nanostructures was observed (**Figure 3b**). In that case, the size of formed nanostructures was greatly reduced. Indeed, minuscule nanoplatelets/nanosheets were formed, and they no longer aggregated forming spherical clusters, but forming layers whose thickness was determined to be 2.1 ± 0.3 μm. The reduced dimensions of these nanosheets might contribute to increase the electrochemically active area, hence resulting in higher electrocatalytic activities of the photoanodes [9].

The differences in shape and size observed for the two synthesized WO₃ nanostructures can be explained by taking into account the interaction between tungsten species and the complexing agent present in the anodization electrolyte (F⁻ or H₂O₂). In order to analyze the formation mechanism of both nanostructures, current density transients obtained during anodization are presented in **Figure 3c**. Both *i* vs *t* curves present common features. At the beginning of the process, current densities abruptly decreased due to the

formation of a compact WO_3 layer as a consequence of substrate oxidation, regardless of the used electrolyte. After a short time, current densities started increasing in both cases due to the dissolution of the previously formed WO_3 layer by the action of F^- and H_2O_2 . Soluble complex tungsten species were formed, according to:

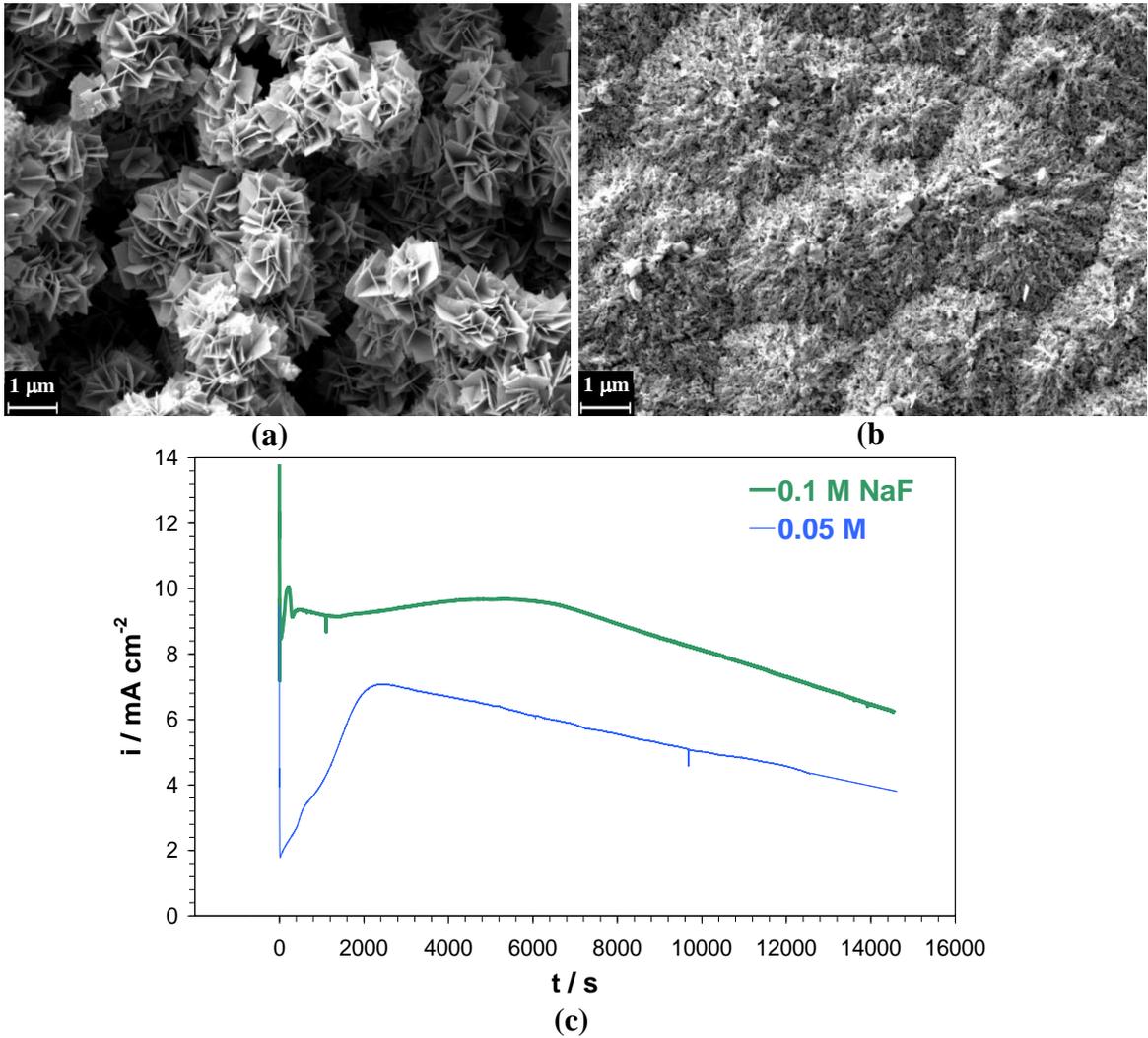
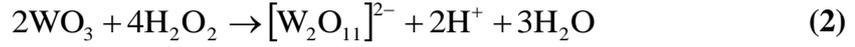
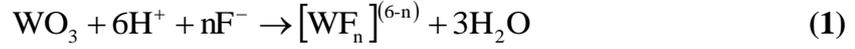


Figure 3. FESEM images of the WO_3 nanostructures synthesized in the 1.5 M H_2SO_4 + 0.1 M NaF electrolyte (a) and in the 1.5 M H_2SO_4 + 0.05 M H_2O_2 electrolyte (b); current density transients obtained during anodization for both WO_3 nanostructures (c).

In both curves, a decrease in current density was also observed after reaching a maximum value. This decrease is directly related to the precipitation of soluble tungsten species formed by reactions (1) and (2) once supersaturation conditions were reached near the electrode surface. This process, characteristic of tungsten compounds, is known as polycondensation [50, 51] and resulted in the formation of the WO_3 nanostructures shown above (**Figure 3a** and **Figure 3b**). Details on the formation mechanism briefly summarized above can be found in a previous study [50].

Nevertheless, there are also visible differences in the curves presented in **Figure 3c**. First, current densities were higher in the NaF electrolyte, which can be explained by the higher concentration of fluorides with respect to H_2O_2 (0.1 M NaF vs 0.05 M H_2O_2). Second, the time at which current density started decreasing (after reaching a maximum value) was shorter in the case of the peroxide-containing solution (~2400 s) than in the F^- solution (~6400 s) (**Figure 3c**). Finally, the slope of the region of decreasing current densities (precipitation of soluble tungsten species) was higher, in absolute value, for the sample fabricated in the NaF solution. Hence, the precipitation of WO_3 nanostructures in the 0.05 M H_2O_2 electrolyte started at shorter times than in the 0.1 M NaF electrolyte, but their rate of formation was slower than in the F^- -containing solution. These facts may explain the morphological and dimensional differences between the two WO_3 nanostructures depicted in **Figures 3a** and **3b**, i.e, nanostructures fabricated in the presence of F^- grew faster than in the presence of H_2O_2 , and hence their distribution was less compact (3D growing) and their size was higher than layer-ordered nanosheets formed in the peroxide-containing solution.

3.2. Photoelectrocatalytic degradation of diuron

Photocurrent densities (i_{ph}) were recorded during the whole degradation experiments and the mean values obtained every 30 min are presented in **Figure 4** for the two WO_3 nanostructures used in this work. Photocurrent densities were always significantly higher for the nanostructures synthesized in the 0.05 M H_2O_2 electrolyte. Photocurrent density values obtained in PEC systems have two main contributions: one coming from the electrocatalytic process and other arising from the photon absorption process. The electrocatalytic process consumes the photogenerated charge carriers and it can become the rate-determining step of the PEC process (the charge transfer at the semiconductor/electrolyte interface) [9]. Current densities in electrochemistry are directly influenced by the real surface active area, since increasing surface area enhances the number of electroactive sites; therefore, from i_{ph} results in **Figure 4**, it can be said that nanosheets obtained in the presence of hydrogen peroxide had a higher electrocatalytic activity than nanoplatelets formed in the F^- -containing electrolyte. Moreover, in the nanostructure formed in the 0.05 M H_2O_2 solution, the very small size of nanosheets and their distribution forming layers, instead of tree-like globular clusters, may result in a reduction of photogenerated holes diffusion path towards the semiconductor/electrolyte interface and may also improve electron movement towards the metallic back contact. Hence, electron-hole recombination would be reduced for nanosheets formed in the peroxide-containing solution and the overall efficiency of the PEC would be enhanced.

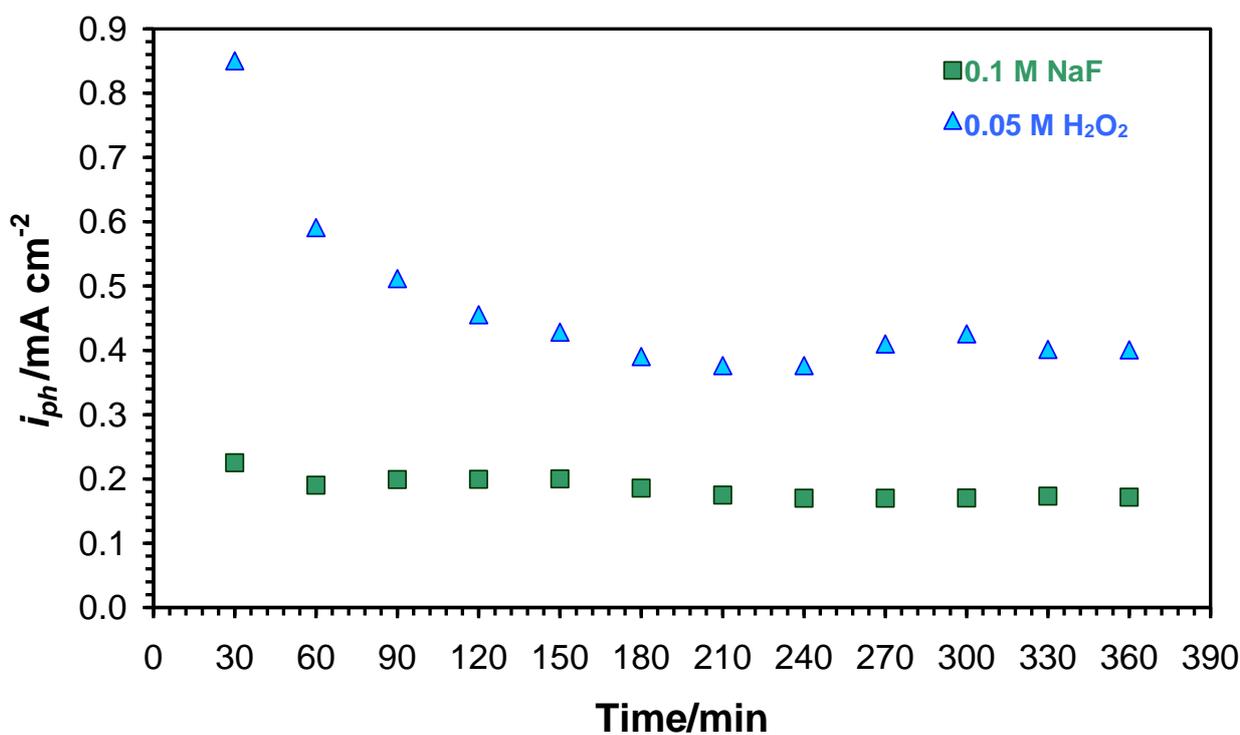


Figure 4. Evolution of photocurrent density (i_{ph}) with degradation time for the two WO₃ nanostructures.

It is worth mentioning that i_{ph} decreased with irradiation time, especially in the nanostructure formed in the 0.05 M H₂O₂ solution. This decrease in photocurrent density has been reported before in the literature for WO₃ photoelectrocatalysts. Several authors have associated this decrease with the formation of H₂O₂ due to incomplete water oxidation at the photoanode surface, followed by the formation of tungsten peroxy-species that are able to partially block the electrochemical active sites [52]. Nevertheless, i_{ph} values reached eventually a high steady-state value, indicating that photoelectrodes were not deactivated. Therefore, the WO₃ samples were photostable under the experimental degradation conditions.

The extent of the degradation reaction was followed by measuring the UV absorbance of the solution with time. **Figure 5** shows the UV absorbance spectra of different

standard solutions of diuron. Concentrations ranged from 1 ppm to 20 ppm, and all the solutions were prepared with 0.1 M H₂SO₄. Two intense bands located at 211 nm and 247 nm can be observed. In general, the band used to monitor the degradation of diuron with time is the second one ($\lambda \approx 247$ nm) [53-60]. This band has been assigned to the typical $\pi \rightarrow \pi^*$ transition of the aromatic ring [60]. The calibration line obtained at 247 nm has also been included in the inset of **Figure 5**.

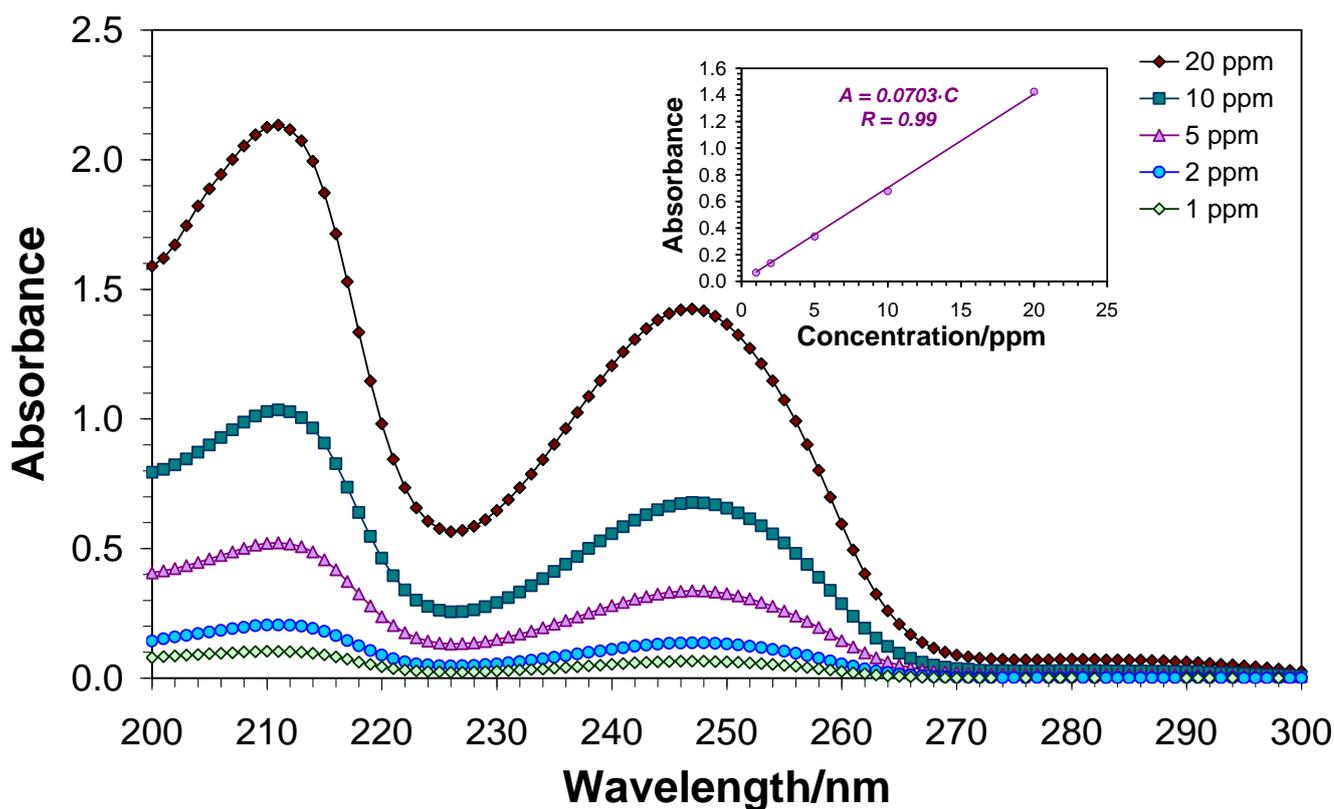


Figure 5. UV absorbance spectra of different solutions of diuron (*inset*: relationship obtained between absorbance A and diuron concentration C at a wavelength of 247 nm).

The evolution of the diuron UV spectra with degradation time is shown in **Figure 6** for the two WO₃ nanostructures used as photoelectrocatalysts in this work. In both cases, absorbance values tended to decrease with reaction time, indicating the elimination of diuron. Neither new absorbance peaks nor displacements of the typical aromatic peak at

247 nm were detected, which may indicate that diuron degradation proceeded through the cleavage of the aromatic ring, instead of proceeding via hydroxylation and chlorine substitution. For example, Papi et al. [59], observed the appearance of an intense absorbance peak at wavelengths of ~300 nm after 4 hours of diuron electrolysis using DSA electrodes, which suggested the formation of a byproduct able to absorb radiation at higher wavelengths than diuron. Kovács et al. [60] observed a slight red shift upon dechlorination of the aromatic ring by gamma radiolysis (from $\lambda_{\text{max}} = 248$ nm for diuron to $\lambda_{\text{max}} = 238$ nm for the non-chlorinated fenuron). Hence, the reduction of the diuron absorbance peaks and the absence of new peaks or shifts in the peak at 247 nm, indicated removal of diuron to form simple molecules [59].

It can also be observed from **Figure 6** that absorbance peaks decreased significantly slower for the WO_3 nanostructure synthesized in the 0.1 M NaF solution (**Figure 6a**) than for the one anodized in the presence of H_2O_2 (**Figure 6b**). Comparison between C/C_0 vs. t plots for both nanostructures is displayed in **Figure 7a**, where C_0 is the initial diuron concentration (20 ppm) and C is diuron concentration at time t obtained from absorbance (A) values at 247 nm using the relationship shown in the inset of **Figure 5**. According to these results, degradation was more efficient using the WO_3 nanosheets obtained in the electrolyte with hydrogen peroxide, which is consistent with the higher photocurrent densities recorded during the degradation process using that photoelectrocatalyst (see **Figure 4**). Degradation efficiency increased from 42 % for nanoplatelets fabricated in the 0.1 M NaF solution to 73 % for nanosheets fabricated in the 0.05 M H_2O_2 solution.

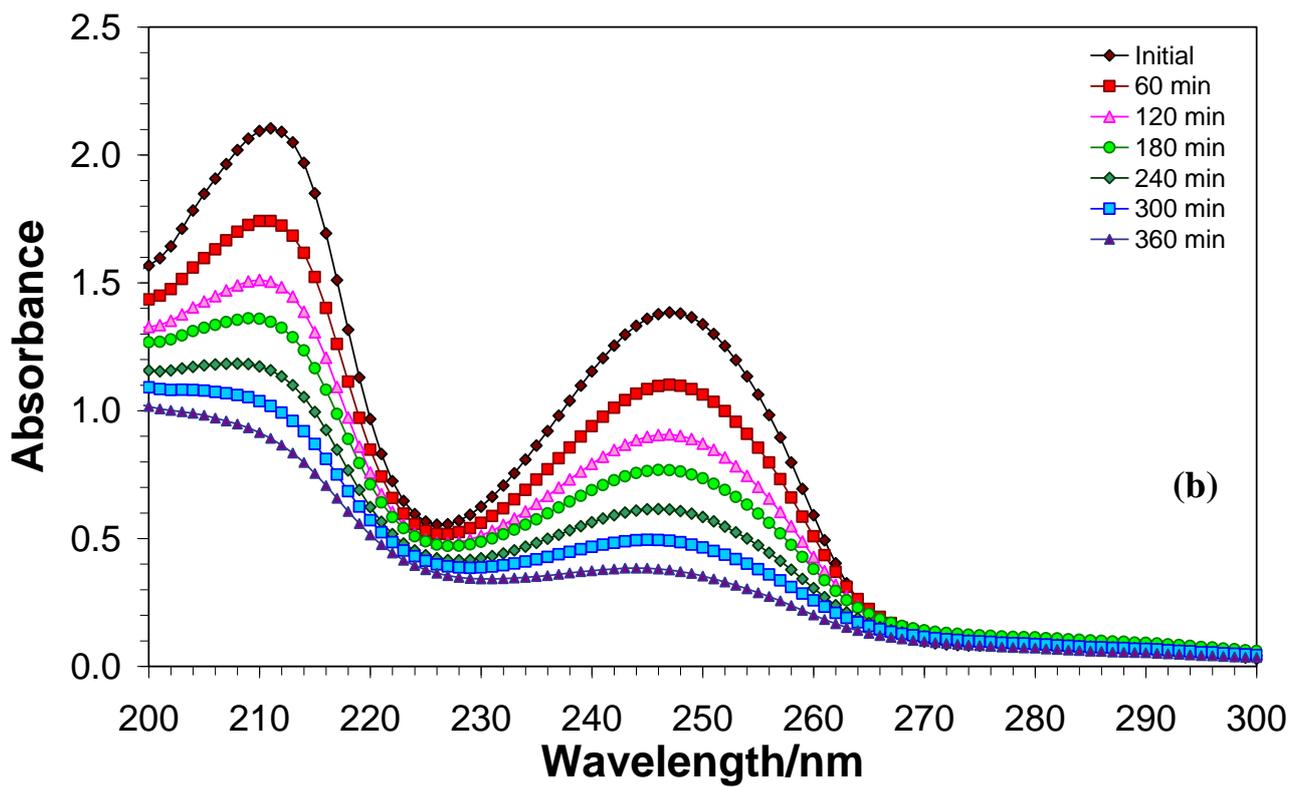
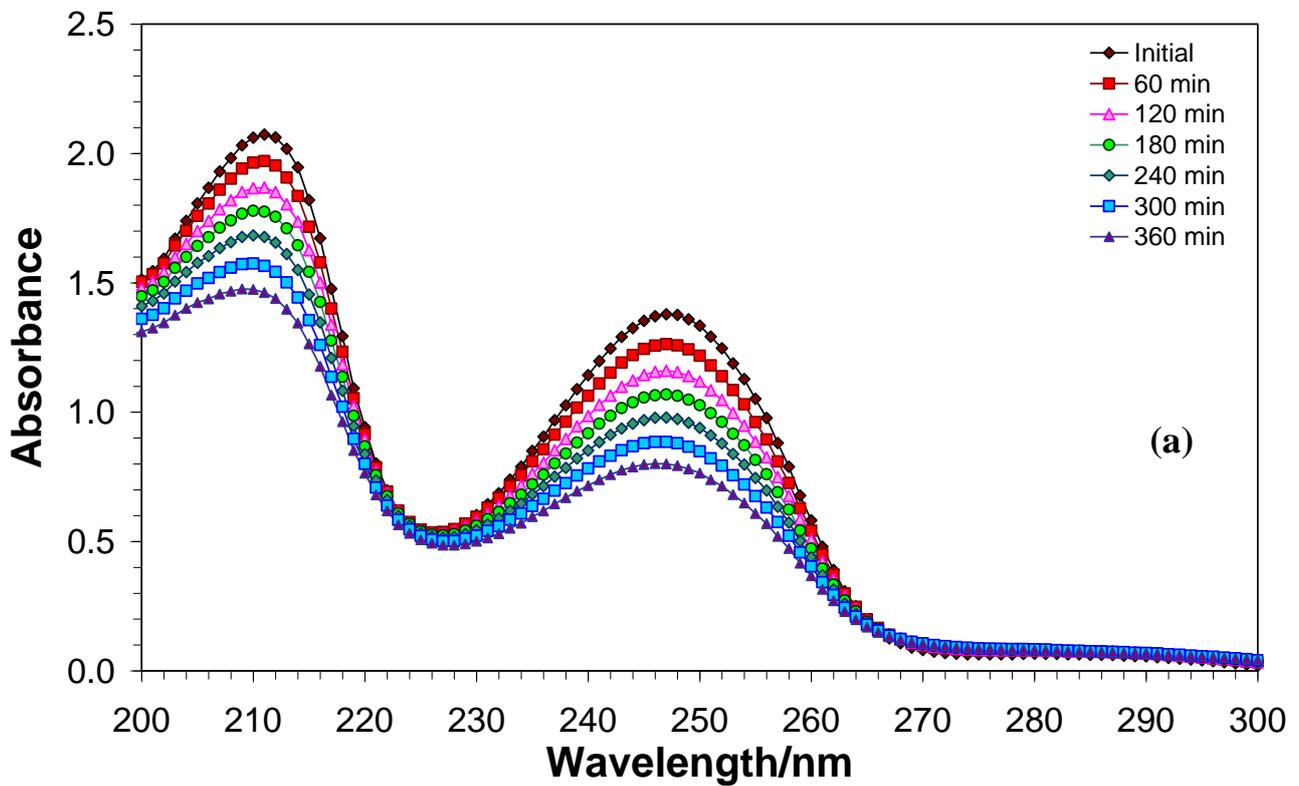


Figure 6. Evolution of the diuron UV spectra with degradation time under visible light (420 nm) for the WO₃ nanostructures anodized in the presence of 0.1 M NaF (a) and 0.05 M H₂O₂ (b).

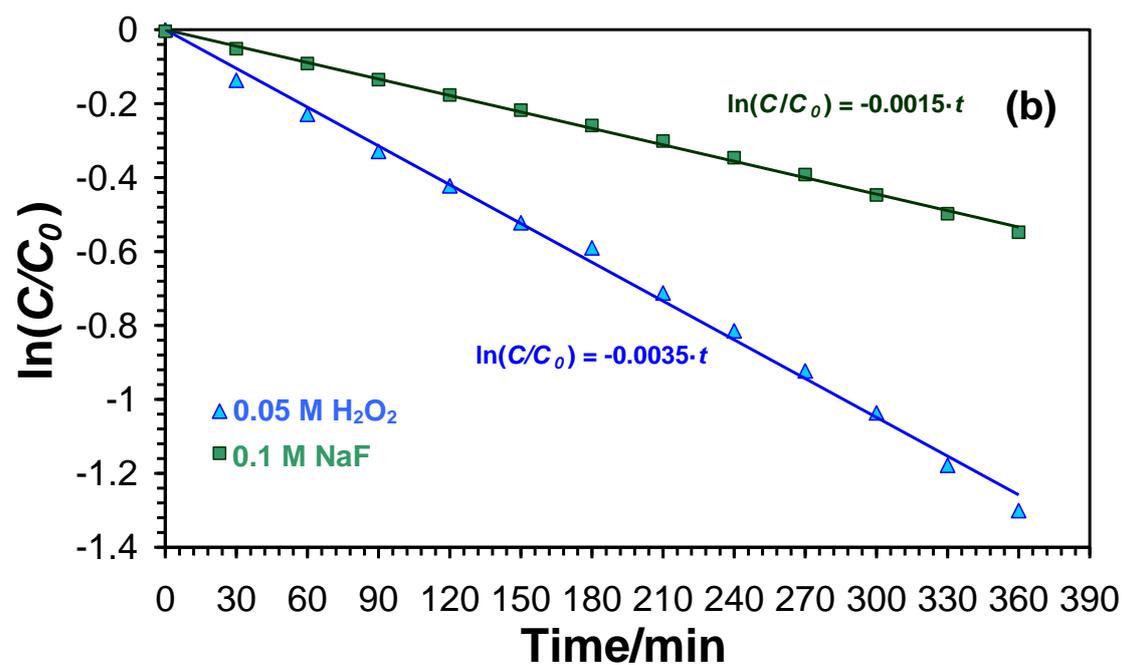
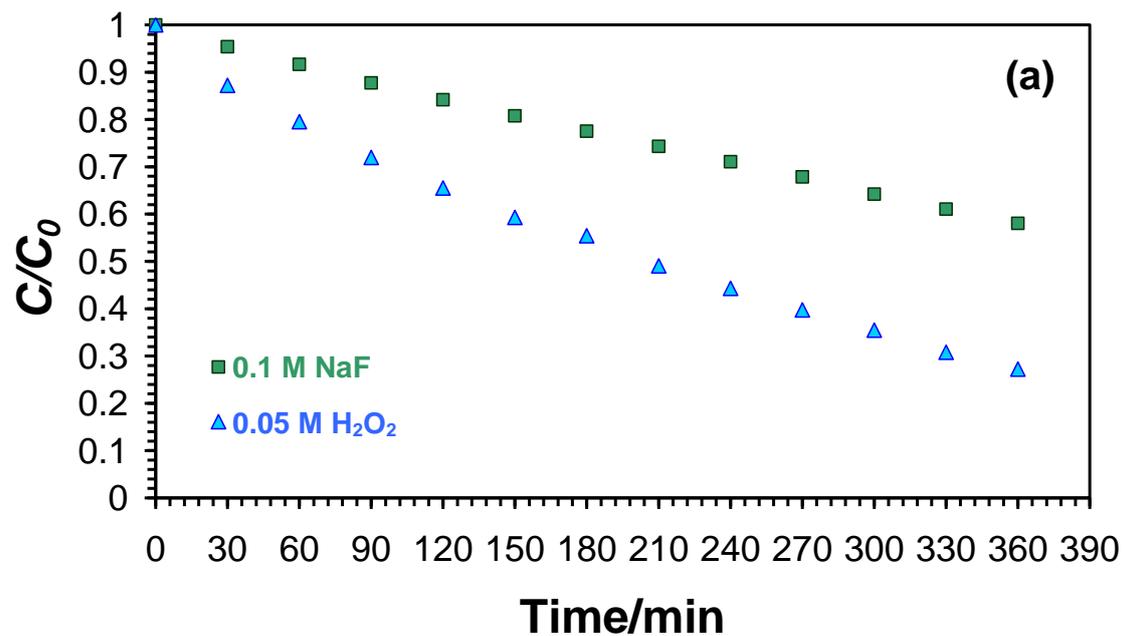


Figure 7. Diuron C/C_0 vs. time plots (a) and $\ln(C/C_0)$ vs. time plots (b) for the two WO_3 nanostructures.

From previous results, and assuming that the PEC degradation of diuron followed pseudo first order kinetics, the kinetic coefficient of the reaction can be determined. According to this model, a plot of $\ln (C/C_0)$ vs. t should give a straight line from whose slope the pseudo first order kinetic coefficient, k' , can be determined (**Figure 7b**). These results are shown in **Table 1** for the two WO_3 nanostructures, together with the coefficients of determination R^2 . Values of R^2 were very close to unity, which indicates the goodness of the linear regression and the validity of the pseudo first order kinetics assumption. An increase of 133 % in k' was obtained by using the WO_3 nanosheets formed in the peroxide-containing electrolyte.

Table 1. Pseudo first order kinetic coefficient k' and coefficients of determination R^2 for diuron degradation using the two WO_3 nanostructures.

Anodization electrolyte	k'/h^{-1}	R^2
0.1 M NaF	0.09	0.998
0.05 M H_2O_2	0.21	0.996

3.3. Ionic chromatography analyses

The concentration of Cl^- and NO_3^- in the solutions after the 360 min of degradation can provide useful information on the extent of diuron mineralization. **Table 2** shows the measured concentrations (in ppm) of these anions in the treated electrolyte for the two WO_3 nanostructures, as well as an indication of the % of mineralization ($\%_{\text{miner}}$), assuming that chlorine atoms present in the diuron molecule were completely

transformed to Cl^- and nitrogen atoms to NO_3^- . Under that hypothesis, the maximum concentration of these anions would be $C_{\text{Cl}^-} = 6.1$ ppm and $C_{\text{NO}_3^-} = 10.6$ ppm.

According to the results presented in **Table 2**, the maximum mineralization degree of diuron was ~50%, using the nanostructure synthesized in the 0.05 M H_2O_2 electrolyte. Notice that these values are all lower than degradation efficiencies shown in **Figure 7** (after 360 min), which indicates that other small molecules were formed from diuron degradation, as expected.

Table 2. Measured concentrations (in ppm) of Cl^- and NO_3^- in the treated diuron electrolyte for the two WO_3 nanostructures. An estimation of the % of mineralization ($\%_{\text{miner}}$) is also shown.

Anodization electrolyte	$C_{\text{Cl}^-}/\text{ppm}$	$\%_{\text{miner}}(\text{Cl}^-)$	$C_{\text{NO}_3^-}/\text{ppm}$	$\%_{\text{miner}}(\text{NO}_3^-)$
0.1 M NaF	1.9	31.1	2.9	27.3
0.05 M H_2O_2	2.9	47.5	5.5	51.9

As it has been mentioned above, mineralization yields shown in **Table 2** were determined supposing a total conversion of $\text{Cl} \rightarrow \text{Cl}^-$ and $\text{N} \rightarrow \text{NO}_3^-$. Yet, the formation of these ions may not be complete. For example, once formed, chloride anions can be oxidized to form gaseous chlorine. On the other hand, the nitrogen in diuron can also be transformed to NH_4^+ cations, apart from forming NO_3^- [61]. Therefore, values shown in **Table 2** should be regarded as *minimum* mineralization yields.

4. CONCLUSIONS

Two different WO_3 nanostructures were obtained by anodization of tungsten in two electrolytes: 1.5 M H_2SO_4 + 0.1 M NaF and 1.5 M H_2SO_4 + 0.05 M H_2O_2 . Nanoplatelets/nanosheets were obtained in both cases, but the dimensions of the nanostructures synthesized in the latter electrolyte were significantly lower due to the interactions between tungsten species and hydrogen peroxide.

Photocurrent densities obtained during visible-light diuron degradation were noticeably higher for the nanostructures fabricated in the peroxide-containing solution. This result arose from the higher electrocatalytic activity (higher surface area) of these nanosheets compared with the ones formed in the F^- -containing electrolyte.

The evolution of diuron absorbance peaks during degradation time indicated removal of the pesticide to form simple molecules, probably through the cleavage of the aromatic ring.

Diuron absorbance peaks decreased significantly faster for the WO_3 nanostructure synthesized in the 0.05 M H_2O_2 solution than for the one anodized in the presence of F^- , which was consistent with the measured photocurrent densities. In fact, the pseudo first order kinetic coefficient for the former nanostructure was 133% higher than for the latter.

A degradation efficiency of 73% and a minimum mineralization degree of 50% were obtained for the nanostructure synthesized in the 0.05 M H₂O₂ electrolyte after 360 min of visible illumination.

Acknowledgements

Authors thank for the financial support to the Ministerio de Economía y Competitividad (Project Code: CTQ2016-79203-R) and for the co-finance by the European Social Fund.

REFERENCES

- [1] H. Zhang, G. Chen, D. W. Bahnemann. Photoelectrocatalytic materials for environmental applications, *J. Mater. Chem.* 19 (2009) 5089-5121.
- [2] R. Dagherir, P. Drogui, D. Robert. Photoelectrocatalytic technologies for environmental applications, *J. Photoch. Photobio. A* 238 (2012) 41-52.
- [3] G. Garcia Bessegato, T. Tasso Guaraldo, M. V. Boldrin Zanoni, Enhancement of Photoelectrocatalysis Efficiency by Using Nanostructured Electrodes, in: M. Aliofkhazraei (ed.), *Modern Electrochemical Methods in Nano, Surface and Corrosion Science*. InTech, London, 2014, pp. 271-319.
- [4] S. Garcia-Segura, E. Brillas. Applied photoelectrocatalysis on the degradation of organic pollutants in wastewaters, *J. Photoch. Photobio. C* 31 (2017) 1-35.
- [5] L.J. Minggu, W.R.W. Daud, M.B. Kassim, An overview of photocells and photoreactors for photoelectrochemical water splitting, *Int. J. Hydrogen Energ.* 35 (2010) 5233-5244.
- [6] R. van de Krol, Principles of Photoelectrochemical Cells, in: R. van de Krol, M. Grätzel (eds.), *Photoelectrochemical Hydrogen Production*, New York, 2012.
- [7] G.G. Bessegato, T.T. Guaraldo, J.F. de Brito, M.F. Brugnera, M.V.B. Zanoni, Achievements and Trends in Photoelectrocatalysis: from Environmental to Energy Applications, *Electrocatal.* 6 (2015) 415-441.
- [8] Q. Zheng, J. Bai, J. Li, L. Li, L. Xia, B. Zhou, Y. Sun, Highly-stable and efficient photocatalytic fuel cell based on an epitaxial TiO₂/WO₃/W nanothorn photoanode and enhanced radical reactions for simultaneous electricity production and wastewater treatment, *Appl. Energ.* 220 (2018) 127-137.

- [9] A. Eftekhari, V. J. Babu, S. Ramakrishna. Photoelectrode nanomaterials for photoelectrochemical water splitting, *Int. J. Hydrogen Energ.* 42 (2017) 11078-11109.
- [10] H. Zheng, J. Z. Ou, M. S. Strano, R. B. Kaner, A. Mitchell, K. Kalantar-zadeh. Nanostructured Tungsten Oxide - Properties, Synthesis, and Applications, *Adv. Funct. Mater.* 21 (2011) 2175-2196.
- [11] J. Luo, M. Hepel. Photoelectrochemical degradation of naphthol blue black diazo dye on WO₃ film electrode, *Electrochim. Acta* 46 (2001) 2913-2922.
- [12] R. Solarska, C. Santato, C. Jorand-Sartoretti, M. Ulmann, J. Augustynski. Photoelectrolytic oxidation of organic species at mesoporous tungsten trioxide film electrodes under visible light illumination, *J. Appl. Electrochem.* 35 (2005) 715-721.
- [13] L. E. Fraga, M. V. Zanoni. Nanoporous of W/WO₃ Thin Film Electrode Grown by Electrochemical Anodization Applied in the Photoelectrocatalytic Oxidation of the Basic Red 51 used in Hair Dye, *J. Braz. Chem. Soc.* 22 (2011) 718-725.
- [14] Q. Zheng, C. Lee. Visible light photoelectrocatalytic degradation of methyl orange using anodized nanoporous WO₃, *Electrochim. Acta* 115 (2014) 140-145.
- [15] S. S. Thind, K. Rozic, F. Amano, A. Chen. Fabrication and photoelectrochemical study of WO₃-based bifunctional electrodes for environmental applications, *Appl. Catal. B Environ* 176-177 (2015) 464-471.
- [16] Y. M. Hunge, V. S. Mohite, S. S. Kumbhar, K. Y. Rajpure, A. V. Moholkar, C. H. Bhosale. Photoelectrocatalytic degradation of methyl red using sprayed WO₃ thin films under visible light irradiation, *J. Mater. Sci.: Mater. Electron.* 26 (2015) 8404-8412.
- [17] S. V. Mohite, V. V. Ganbavle, K. Y. Rajpure. Solar photoelectrocatalytic activities of rhodamine-B using sprayed WO₃ photoelectrode, *J. Alloy. Compd.* 655 (2016) 106-113.
- [18] Y. M. Hunge, M. A. Mahadik, S. S. Kumbhar, V. S. Mohite, K. Y. Rajpure, N. G. Deshpande, A. V. Moholkar, C. H. Bhosale. Visible light catalysis of methyl orange using nanostructured WO₃ thin films, *Ceram. Int.* 42 (2016) 789-798.
- [19] J. Zheng, J. Li, J. Bai, X. Tan, Q. Zeng, L. Li, B. Zhou. Efficient Degradation of Refractory Organics Using Sulfate Radicals Generated Directly from WO₃ Photoelectrode and the Catalytic Reaction of Sulfate, *Catalysts* 7 (2017) 346.
- [20] S. V. Mohite, V. V. Ganbavle, K. Y. Rajpure. Photoelectrocatalytic activity of immobilized Yb doped WO₃ photocatalyst for degradation of methyl orange dye, *J. Energy Chem.* 26 (2017) 440-447.

- [21] R. M. Fernández-Domene, R. Sánchez-Tovar, B. Lucas-Granados, C. S. García-Zamora, J. García-Antón. Customized WO₃ nanoplatelets as visible-light photoelectrocatalyst for the degradation of a recalcitrant model organic compound (methyl orange), *J. Photochem. Photobiol. A: Chem.* 356 (2018) 46-56.
- [22] E. O. Scott-Emuakpor, A. Kruth, M. J. Todd, A. Raab, G. I. Paton, D. E. Macphee. Remediation of 2,4-dichlorophenol contaminated water by visible light-enhanced WO₃ photoelectrocatalysis, *Appl. Catal. B: Environ.* 123-124 (2012) 433-439.
- [23] S. S. Thind, M. Tian, A. Chen. Direct growth and photo-electrochemical study of WO₃ nanostructured materials, *Electrochem. Commun.* 43 (2014) 13-17.
- [24] S. Xie, K. Ouyang. Degradation of refractory organic compounds by photocatalytic fuel cell with solar responsive WO₃/FTO photoanode and air-breathing cathode, *J. Colloid Interf. Sci.* 500 (2017) 220-227.
- [25] G. Longobucco, L. Pasti, A. Molinari, N. Marchetti, S. Caramori, V. Cristino, R. Boaretto, C. A. Bignozzi. Photoelectrochemical mineralization of emerging contaminants at porous WO₃ interfaces, *Appl. Catal. B: Environ.* 204 (2017) 273-282.
- [26] S. V. Mohite, V. V. Ganbavle, K. Y. Rajpure. Photoelectrochemical performance and photoelectrocatalytic degradation of organic compounds using Ga:WO₃ thin films, *J. Photochem. Photobiol. A: Chem.* 344 (2017) 56-63.
- [27] Q. Zeng, J. Li, L. Li, J. Bai, L. Xia, B. Zhou. Synthesis of WO₃/BiVO₄ photoanode using a reaction of bismuth nitrate with peroxovanadate on WO₃ film for efficient photoelectrocatalytic water splitting and organic pollutant degradation, *Appl. Catal. B: Environ.* 217 (2017) 21-29.
- [28] Diuron Information - European Chemicals Agency (ECHA) (<https://echa.europa.eu/substance-information/-/substanceinfo/100.005.778>). (Accessed 8 May 2018).
- [29] Environmental Risk Assessment for the Reregistration of Diuron - United States Environmental Protection Agency (EPA), 2001.
- [30] F. Akcha, A. Barranger, E. Bachère, C. H. Berthelin, D. Piquemal, P. Alonso, R. R. Sallan, G. Dimastrogiovanni, C. Porte, D. Menard, A. Szczybelski, A. Benabdelmouna, M. Auffret, J. Rouxel, T. Burgeot. Effects of an environmentally relevant concentration of diuron on oyster genitors during gametogenesis: responses of early molecular and cellular markers and physiological impacts, *Environ. Sci. Pollut. Res.* 23 (2016) 8008-8020.
- [31] T. S. Boscolo Pereira, C. N. Pereira Boscolo, A. A. Felício, S. R. Batlouni, D. Schlenk, E. ves de Almeida. Estrogenic activities of diuron metabolites in female Nile tilapia (*Oreochromis niloticus*), *Chemosphere* 146 (2016) 497-502.

- [32] M. Velki, C. Di Paolo, J. Nelles, T. B. Seiler, H. Hollert. Diuron and diazinon alter the behavior of zebrafish embryos and larvae in the absence of acute toxicity, *Chemosphere* 180 (2017) 65-76.
- [33] M. R. M. Khanam, Y. Shimasaki, M. Z. Hosain, K. Mukai, M. Tsuyama, X. Qiu, R. Tasmin, H. Goto, Y. Oshima. Diuron causes sinking retardation and physiochemical alteration in marine diatoms *Thalassiosira pseudonana* and *Skeletonema marinoi-dohrnii* complex, *Chemosphere* 175 (2017) 200-209.
- [34] S. Hasenbein, J. Peralta, S. P. Lawler, R. E. Connon. Environmentally relevant concentrations of herbicides impact non-target species at multiple sublethal endpoints, *Sci. Total Environ.* 607-608 (2017) 733-743.
- [35] D. J. Rowen, M. A. Templeman, M. J. Kingsford. Herbicide effects on the growth and photosynthetic efficiency of *Cassiopea maremetens*, *Chemosphere* 182 (2017) 143-148.
- [36] C. N. P. Boscolo, T. S. B. Pereira, I. G. Batalhão, P. L. R. Dourado, D. Schlenk, E. A. de Almeida. Diuron metabolites act as endocrine disruptors and alter aggressive behavior in Nile tilapia (*Oreochromis niloticus*), *Chemosphere* 191 (2018) 832-838.
- [37] S. Beernaerts, M. Gerard, P. Debongnie, J. P. Barthelemy, L. Pussemier, A. Copin. Two Years of Pesticides Monitoring in a Belgian Watershed, *Int. J. Environ. An. Ch.* 83 (2003) 469-480.
- [38] R. Loos, R. Carvalho, D. C. António, S. Comero, G. Locoro, S. Tavazzi, B. Paracchini, M. Ghiani, T. Lettieri, L. Blaha, B. Jarosova, S. Voorspoels, K. Servaes, P. Haglund, J. Fick, R. H. Lindberg, D. Schwesig, B. M. Gawlik. EU-wide monitoring survey on emerging polar organic contaminants in wastewater treatment plant effluents, *Water Res.* 47 (2013) 6475-6487.
- [39] T. D. H. Le, A. Scharmüller, M. Kattwinkel, R. Kühne, G. Schüürmann, R. B. Schäfer. Contribution of waste water treatment plants to pesticide toxicity in agriculture catchments, *Ecotox. Environ. Safe.* 145 (2017) 135-141.
- [40] Z. Tousova, P. Oswald, J. Slobodnik, L. Blaha, M. Muz, M. Hu, W. Brack, M. Krauss, C. Di Paolo, Z. Tarcai, T. B. Seiler, H. Hollert, S. Koprivica, M. Ahel, J. E. Schollée, J. Hollender, M. J. F. Suter, A. O. Hidasi, K. Schirmer, M. Sonavane, S. it-Aissa, N. Creusot, F. Brion, J. Froment, A. C. Almeida, K. Thomas, K. E. Tollefsen, S. Tufi, X. Ouyang, P. Leonards, M. Lamoree, V. O. Torrens, A. Kolkman, M. Schriks, P. Spirhanzlova, A. Tindall, T. Schulze. European demonstration program on the effect-based and chemical identification and monitoring of organic pollutants in European surface waters, *Sci. Total Environ.* 601-602 (2017) 1849-1868.
- [41] J. Yuan, M. I. Van Dyke, P. M. Huck. Identification of critical contaminants in wastewater effluent for managed aquifer recharge, *Chemosphere* 172 (2017) 294-301.

- [42] H. Chen, Y. Luo, C. Potter, P. J. Moran, M. L. Grieneisen, M. Zhang. Modeling pesticide diuron loading from the San Joaquin watershed into the Sacramento-San Joaquin Delta using SWAT, *Water Res.* 121 (2017) 374-385.
- [43] B. Reoyo-Prats, D. Aubert, C. Menniti, W. Ludwig, J. Sola, M. Pujo-Pay, P. Conan, O. Verneau, C. Palacios. Multicontamination phenomena occur more often than expected in Mediterranean coastal watercourses: Study case of the Têt River (France), *Sci. Total Environ.* 579 (2017) 10-21.
- [44] M. A. Rippey, A. Deletic, J. Black, R. Aryal, J. L. Lampard, J. Y.-M. Tang, D. McCarthy, P. Kolotelo, J. Sidhu, W. Gernjak. Pesticide occurrence and spatio-temporal variability in urban run-off across Australia, *Water Res.* 115 (2017) 245-255.
- [45] N. A. Munz, F. J. Burdon, D. de Zwart, M. Junghans, L. Melo, M. Reyes, U. Schönenberger, H. P. Singer, B. Spycher, J. Hollender, C. Stamm. Pesticides drive risk of micropollutants in wastewater-impacted streams during low flow conditions, *Water Res.* 110 (2017) 366-377.
- [46] R. Münze, C. Hannemann, P. Orlinskiy, R. Gunold, A. Paschke, K. Foit, J. Becker, O. Kaske, E. Paulsson, M. Peterson, H. Jernstedt, J. Kreuger, G. Schüürmann, M. Liess. Pesticides from wastewater treatment plant effluents affect invertebrate communities, *Sci. Total Environ.* 599-600 (2017) 387-399.
- [46] L. Le Coadou, K. Le Ménach, P. Labadie, M. H. Dévier, P. Pardon, S. Augagneur, H. Budzinski. Quality survey of natural mineral water and spring water sold in France: Monitoring of hormones, pharmaceuticals, pesticides, perfluoroalkyl substances, phthalates, and alkylphenols at the ultra-trace level, *Sci. Total Environ.* 603-604 (2017) 651-662.
- [48] E. Herrero-Hernández, M. S. Rodríguez-Cruz, E. Pose-Juan, S. Sánchez-González, M. S. Andrades, M. J. Sánchez-Martín. Seasonal distribution of herbicide and insecticide residues in the water resources of the vineyard region of La Rioja (Spain), *Sci. Total Environ.* 609 (2017) 161-171.
- [49] L. H. Nowell, P. W. Moran, T. S. Schmidt, J. E. Norman, N. Nakagaki, M. E. Shoda, B. J. Mahler, P. C. Van Metre, W. W. Stone, M. W. Sandstrom, M. L. Hladik. Complex mixtures of dissolved pesticides show potential aquatic toxicity in a synoptic study of Midwestern U.S. streams, *Sci. Total Environ.* 613-614 (2018) 1469-1488.
- [50] R. M. Fernández-Domene, R. Sánchez-Tovar, B. Lucas-Granados, G. Roselló-Márquez, J. García-Antón. A simple method to fabricate high-performance nanostructured WO₃ photocatalysts with adjusted morphology in the presence of complexing agents, *Mater. Design* 116 (2017) 160-170.
- [51] E. Lassner and W. D. Schubert, Tungsten. Properties, Chemistry, Technology of the Element, Alloys, and Chemical Compounds, Kluwer Academic / Plenum Publishers, New York, 1999.

- [52] J. A. Seabold, K. S. Choi. Effect of a Cobalt-Based Oxygen Evolution Catalyst on the Stability and the Selectivity of Photo-Oxidation Reactions of a WO_3 Photoanode, *Chem. Mater.* 23 (2011) 1105-1112.
- [53] M. Van Boven, L. Laruelle, P. Daenens. HPLC Analysis of Diuron and Metabolites in Blood and Urine, *J. Anal. Toxicol.* 14 (1990) 231-234.
- [54] P. Mazellier, J. Jirkovsky, M. Bolte. Degradation of Diuron Photoinduced by Iron(III) in Aqueous Solution, *Pestic. Sci.* 49 (1997) 259-267.
- [55] J. Jirkovský, V. Faure, P. Boule. Photolysis of Diuron, *Pestic. Sci.* 50 (1999) 42-52.
- [56] K. Djebbar, T. Sehili, P. Mazellier, J. De Laat. Phototransformation of diuron in aqueous solution by UV irradiation in the absence and in the presence of H_2O_2 , *Environ. Technol.* 24 (2003) 479-489.
- [57] M. J. Farré, M. I. Maldonado, W. Gernjak, I. Oller, S. Malato, X. Domènech, J. Peral. Coupled solar photo-Fenton and biological treatment for the degradation of diuron and linuron herbicides at pilot scale, *Chemosphere* 72 (2008) 622-629.
- [58] F. Galindo-Hernández, J. A. Wang, R. Gómez, X. Bokhimi, L. Lartundo, A. Mantilla. Structural modifications in $\text{Au}/\text{Al}_2\text{O}_3\text{-CeO}_2$ mixed oxides as a function of Ce^{4+} content and its effects in the mineralization of the herbicide diuron, *J. Photochem. Photobiol. A: Chem.* 243 (2012) 23-32.
- [59] A. R. F. Pippi, S. A. Neto, A. R. De Andrade. Electrochemical Degradation of Diuron in Chloride Medium using DSA® Based Anodes, *J. Braz. Chem. Soc.* 24 (2013) 1259-1266.
- [60] K. Kovács, S. He, V. Mile, T. Csay, E. Takács, L. Wojnárovits. Ionizing radiation induced degradation of diuron in dilute aqueous solution, *Chem. Cent. J.* 9 (2015) 21.
- [61] D. Bamba, P. Atheba, D. Robert, A. Trokourey, B. Dongui. Photocatalytic degradation of the diuron pesticide, *Environ. Chem. Lett.* 6 (2008) 163-167.